

Evaluating the Role of Nitrification in Pharmaceutical Biodegradation

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Microconstituents & Emerging Contaminants

- Endocrine disrupting compounds
- Pharmaceuticals
- Personal care products
- Persistent Organic Pollutants
- Prions
- Nanoparticles



Focus on Pharmaceuticals (PhACs)



motivation for research

- PhACs (in general) are important in modern society – clear benefits to human and animal health
- PhACs are complex molecules that are often *designed* to be bioactive
- Research has only begun to explore the influence of PhACs in the natural and engineered environments (maybe 10% of most used PhACs have been studied; even less have been extensively studied)
- Increasing evidence to suggest PhACs can have deleterious environmental effects.
- Influence and effects of long-term low-concentration (\ll therapeutic doses) *mixtures* on ecosystem health is an *open question*.
- Current evidence suggests some PhACs may be attenuated more in biological treatment systems employed for nutrient management

Pharmaceuticals

low concentration, long-term exposure – mixture effects?

Research over the last decade or so, has identified:

- Presence of pharmaceuticals in our waters (landmark study: Kolpin et al. 2002 *Environ Sci. Technol.*)
- Certain pharmaceuticals can alter behavior, mobility and development in aquatic species (e.g., Ruiz et. al., 2010 *Environ. Sci. Techol.*; Quinn et. al., 2009, *Sci Total Environ*; Fent et. al., 2006, *Aquat. Toxicol.*)



More recently, there is preliminary evidence to suggest pharmaceuticals may:

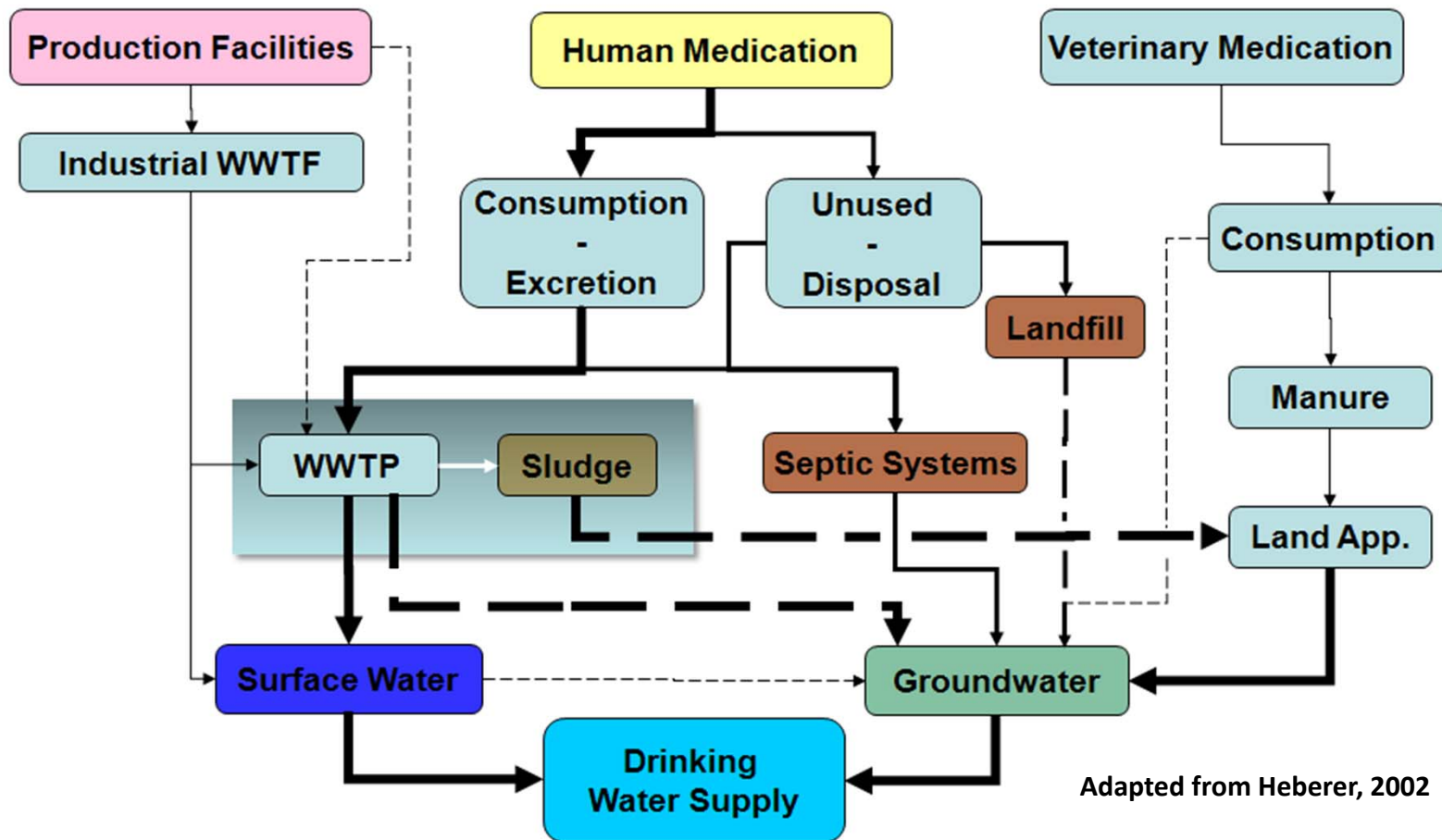
- Bioaccumulate in plants and animals (e.g., Zenker et al. 2014 review *J. Environ. Mgt.*; Wu et al. 2014, *Environ. Sci. Technol.*)
- Cause change in animal behavior and physiology outside of the aquatic environment (e.g., birds - Bean et al. 2014 *Phil. Trans. R. Soc. B*)

Links below are videos describing *Phil. Trans. R. Soc. B.* edition (as well as the starling study referenced above) related to PhAC impacts on wildlife and ecosystems

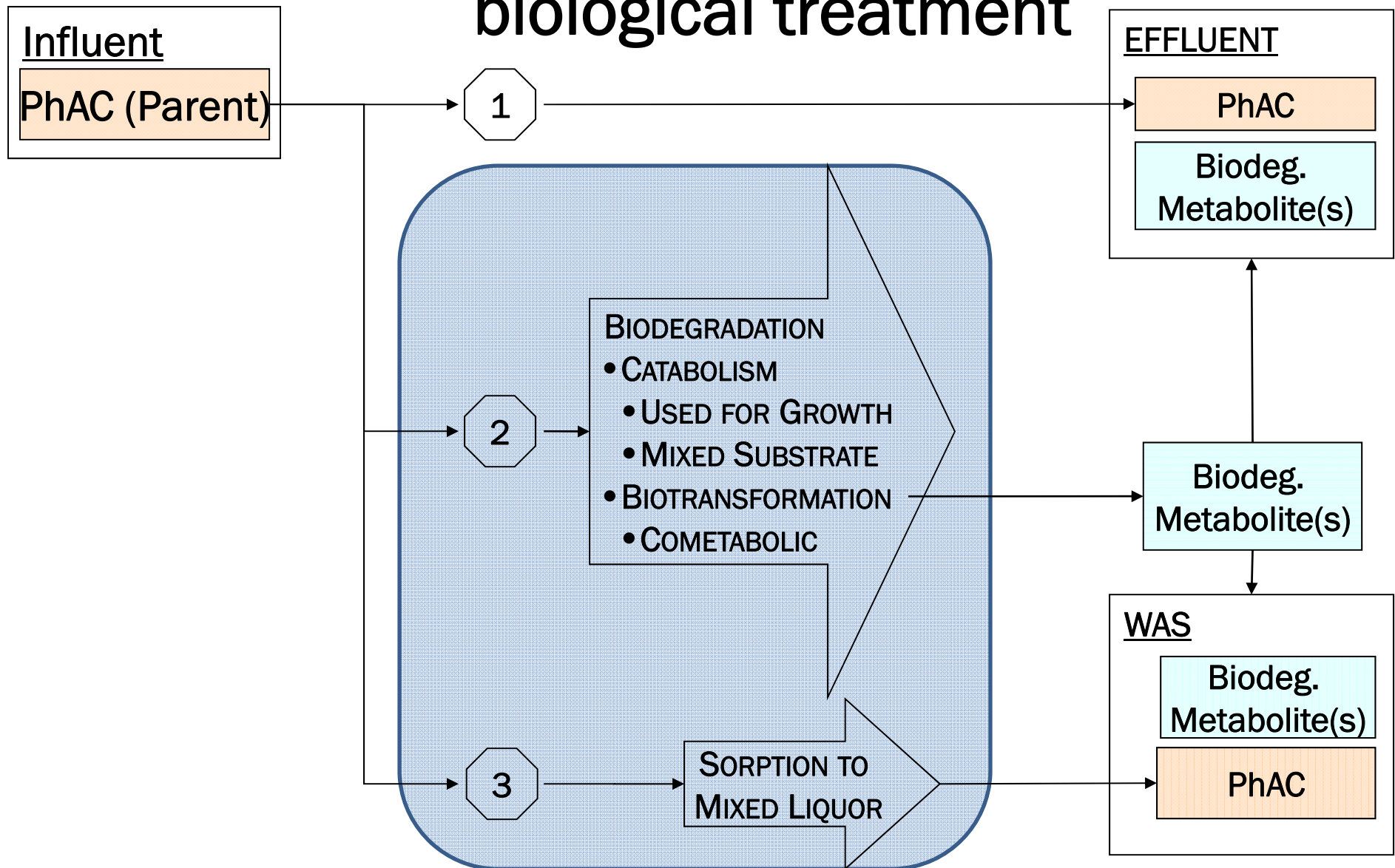
http://www.youtube.com/watch?v=oByacXyaH00&feature=player_embedded

http://www.youtube.com/watch?feature=player_embedded&v=HxBub1IIIE

Pharmaceuticals in the Environment



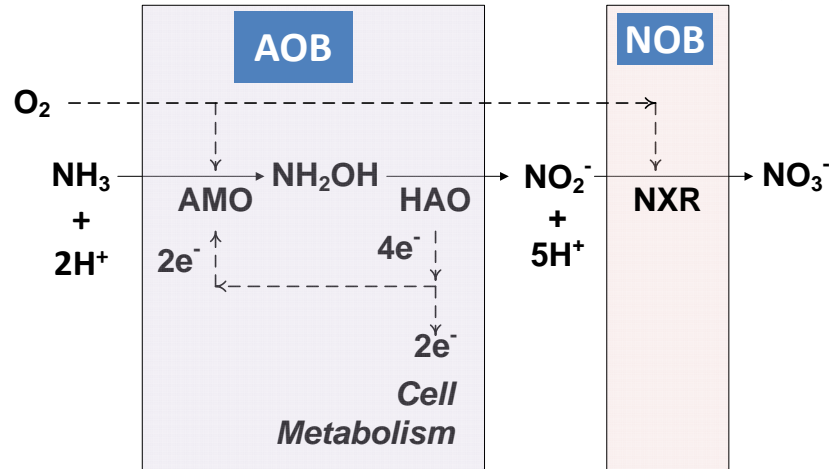
Pharmaceutical attenuation: biological treatment



Focus on Nitrification (denitrification in future)

acknowledged that heterotrophs have important role

Ammonia Oxidizing Bacteria Nitrite Oxidizing Bacteria



- Nitrification is an essential process to manage the nitrogen cycle - NAE grand challenge – (NAE, 2008)

- WWTPs are required to meet increasingly stringent effluent nitrogen criteria

- Ammonia Monooxygenase (AMO) has broad substrate range (Keener and Arp, 1994; Hooper et. al., 1997; Skotnicka-Pitak et. al., 2009, Taher and Chandra, 2013)

Figure adapted from Arp & Stein (2003)

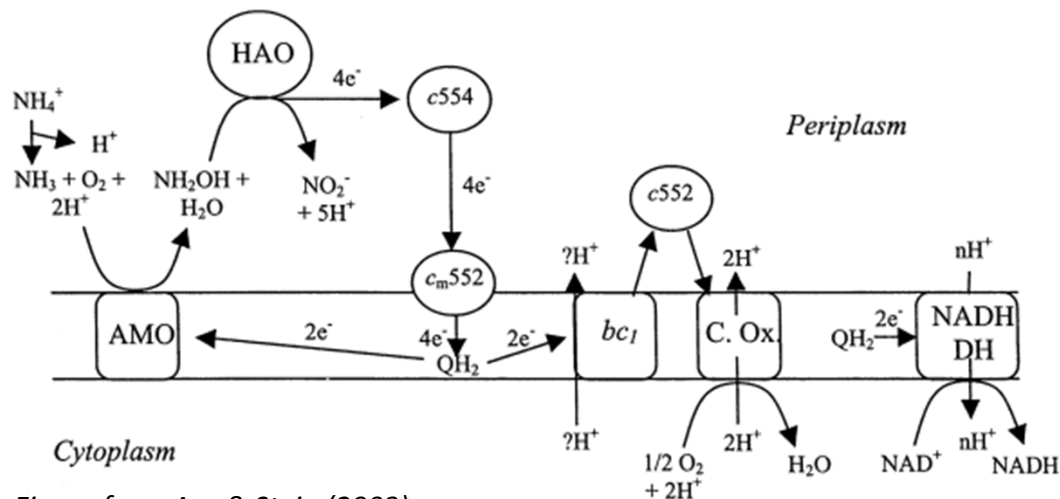
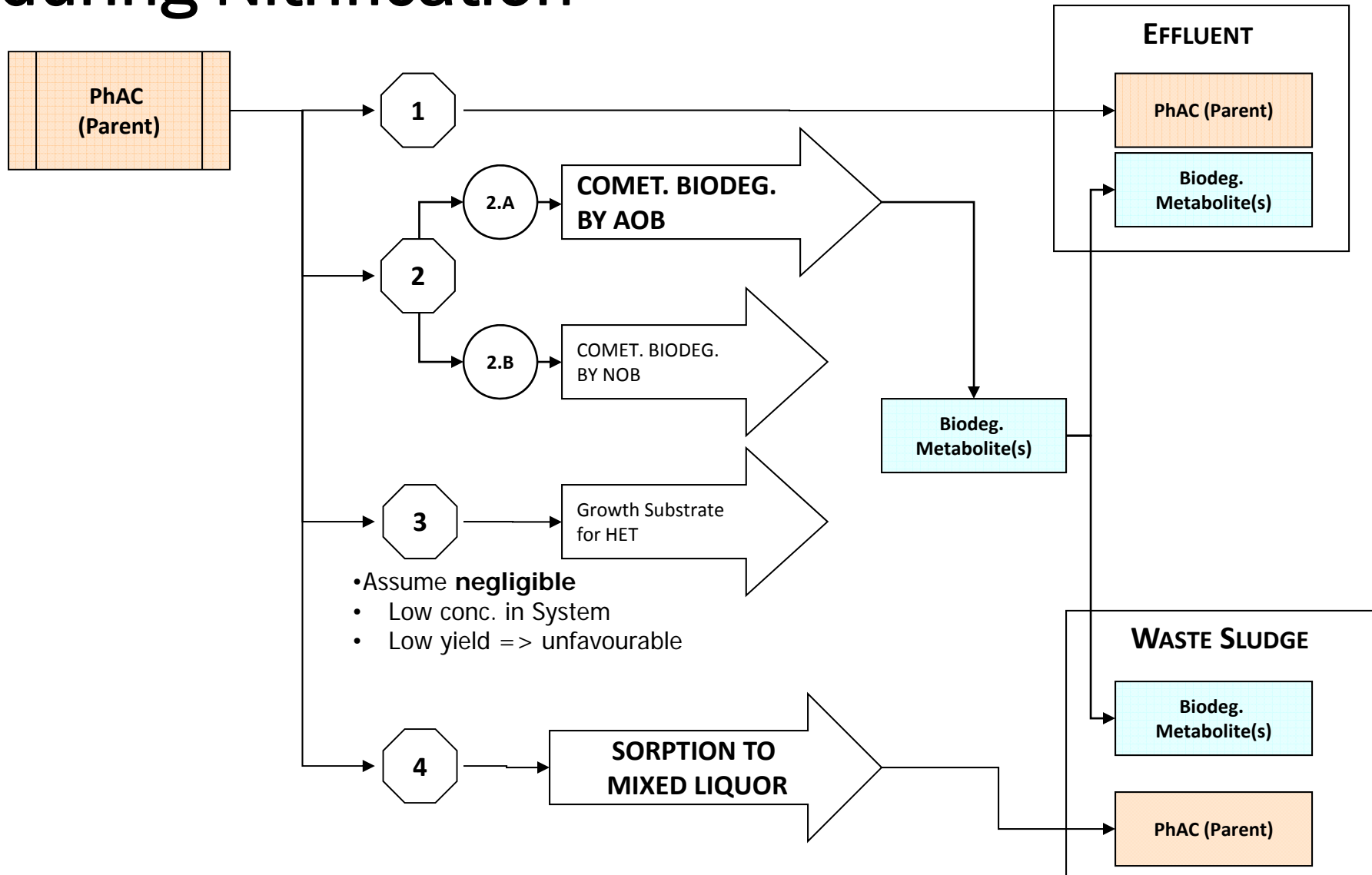


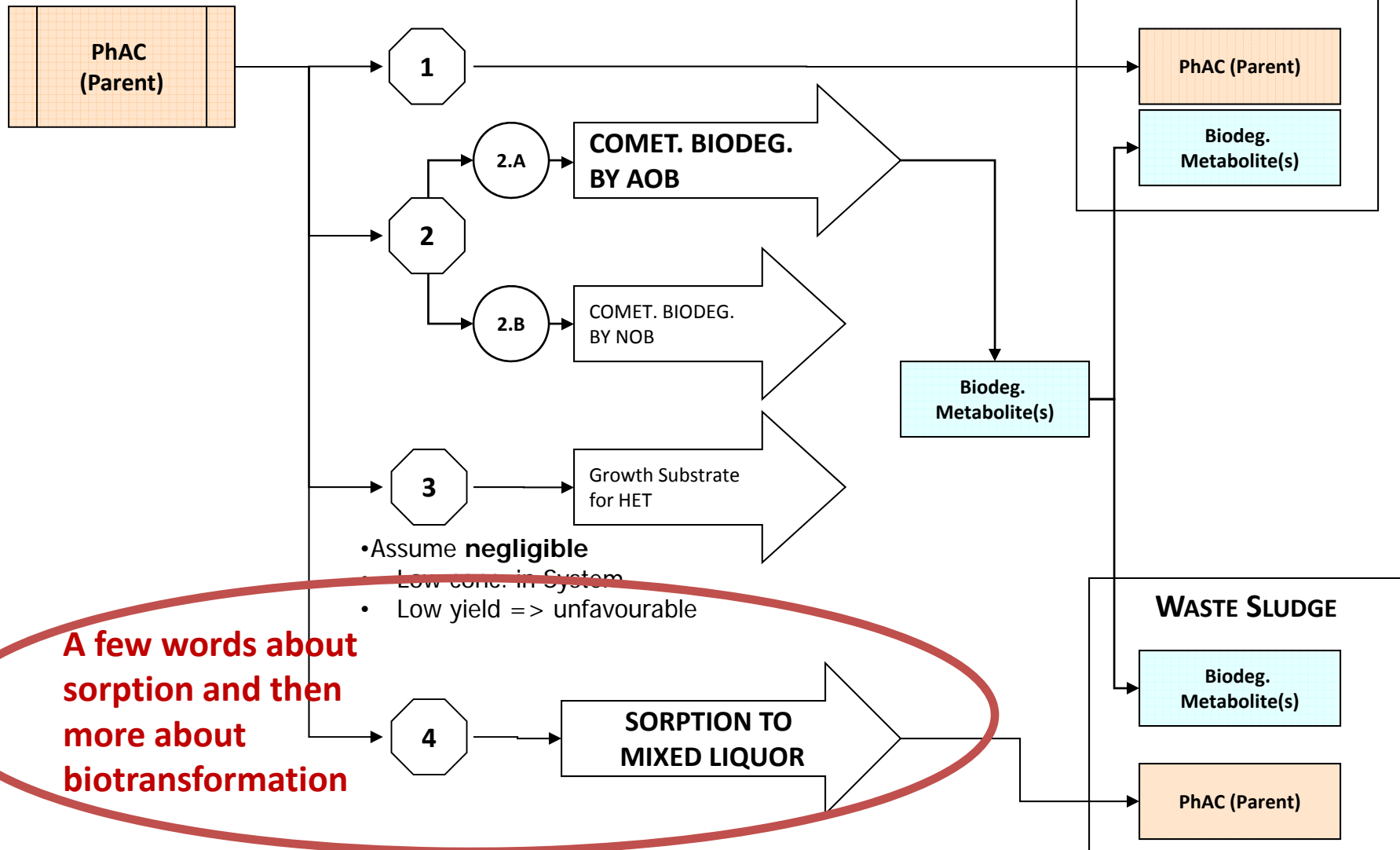
Figure from Arp & Stein (2003)

- AMO: Ammonia Monooxygenase (membrane bound protein)
- HAO: Hydroxylamine Oxidoreductase (periplasmic protein)
- NXR: Nitrite Oxidoreductase

Pharmaceutical Attenuation during Nitrification

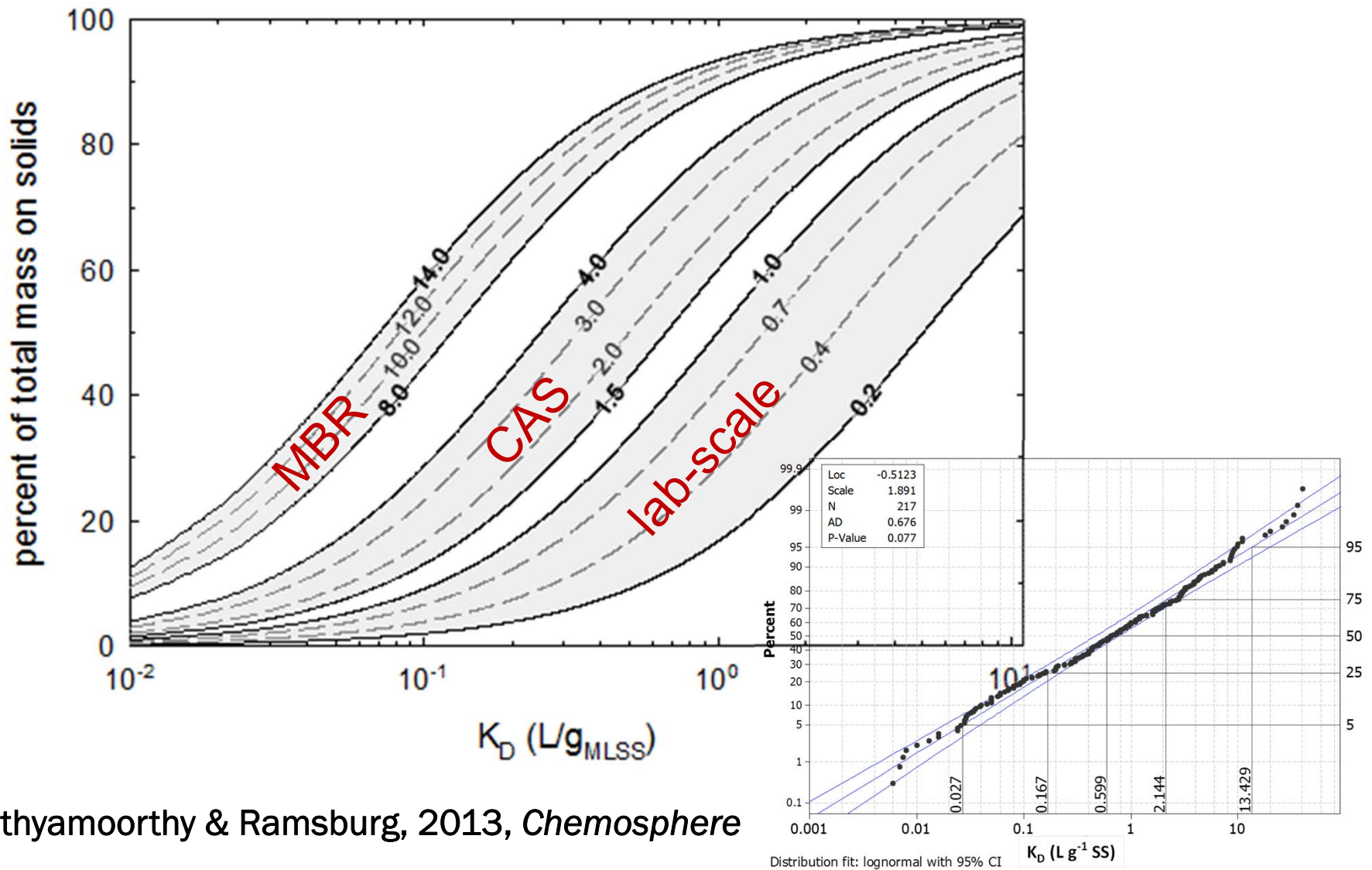


Pharmaceutical Attenuation during Nitrification



A few words about sorption and then more about biotransformation

Possible role of sorption & how to predict it



Sathyamoorthy & Ramsburg, 2013, *Chemosphere*

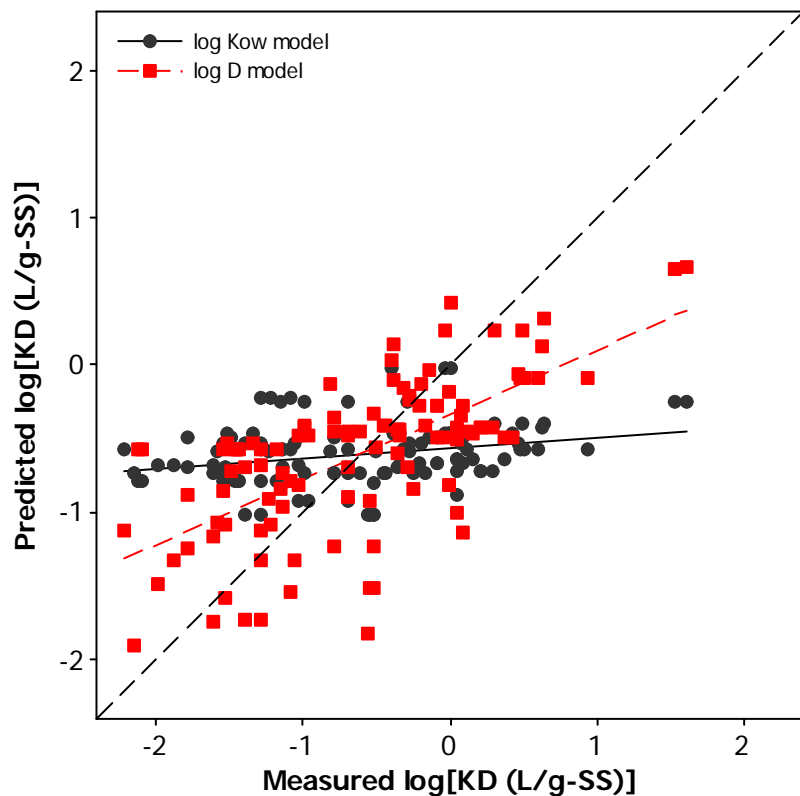
K_{ow} and D models

sorption to activated sludge solids



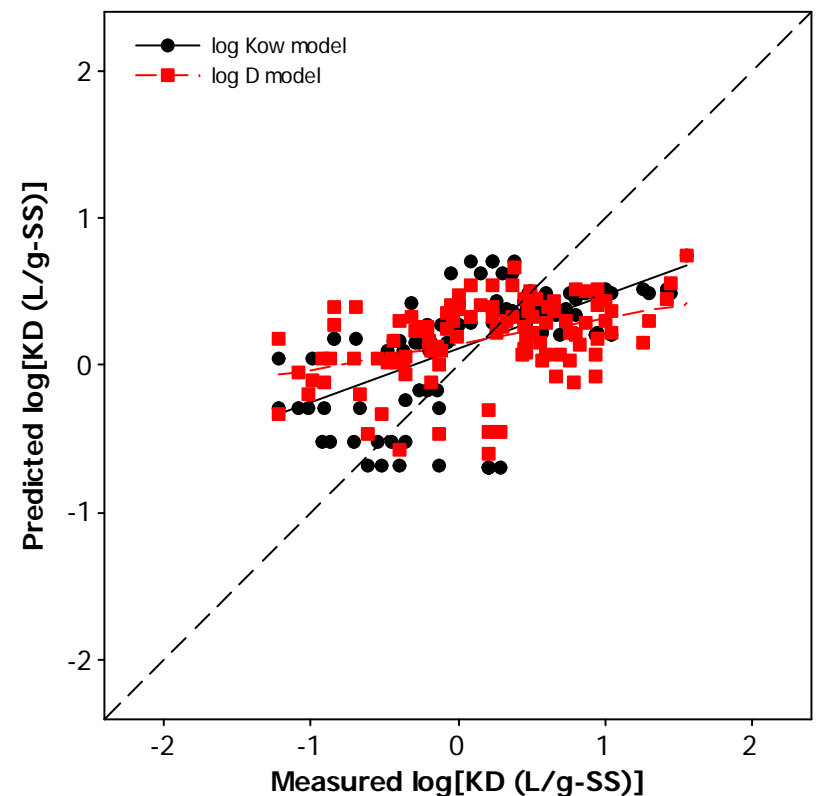
Negatively Charged and Uncharged

Negatively Charged & Neutral Data ($n_{DATA} = 109$, $n_{PHACs} = 30$)							
Model Summary			Model Performance				
Predictor	Coeff.	SE.Coeff.	S	R ²	adj-R ²	pred-R ²	NSE
Constant	-1.151	0.201	0.77	7.2%	6.3%	3.9%	0.07
log K_{ow}	0.144	0.050					
Constant	-1.379	0.101	0.60	44.0%	43.5%	41.8%	0.44
log D	0.327	0.036					



Positively Charged

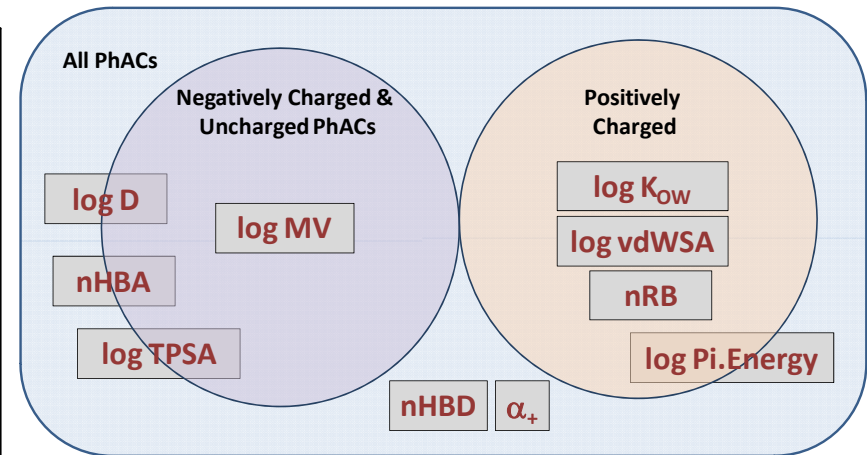
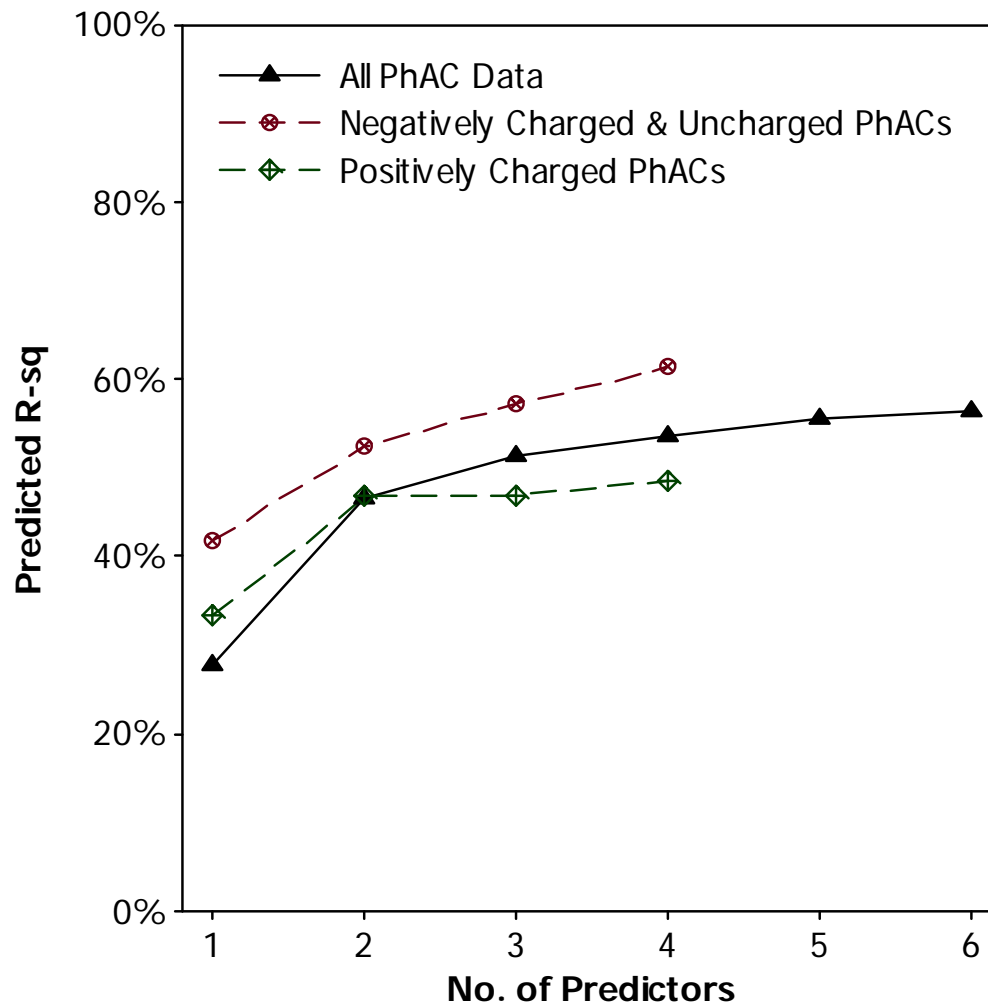
Positively Charged Data ($n_{DATA} = 108$, $n_{PHACs} = 32$)							
Model Summary			Model Performance				
Predictor	Coeff.	SE.Coeff.	S	R ²	adj-R ²	pred-R ²	NSE
Constant	-0.738	0.128	0.51	36.1%	35.5%	33.6%	0.36
log K_{ow}	0.237	0.031					
Constant	-0.108	0.082	0.58	17.4%	16.6%	14.2%	0.17
log D	0.144	0.030					



Sathyamoorthy & Ramsburg, 2013, *Chemosphere*

Polyparameter models

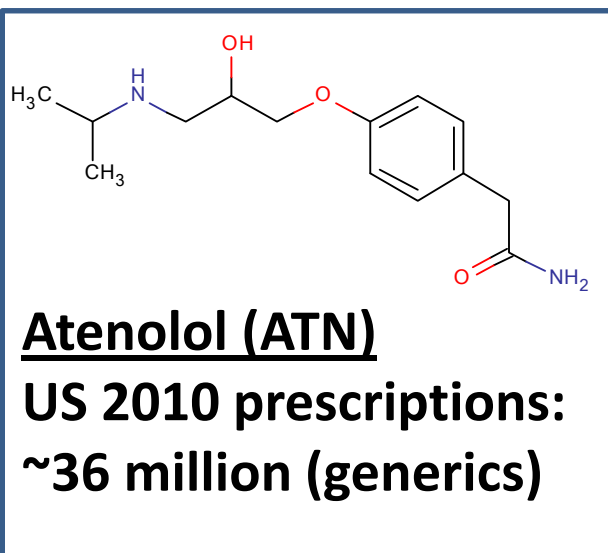
improved predictive capability for sorption to activate sludge solids?



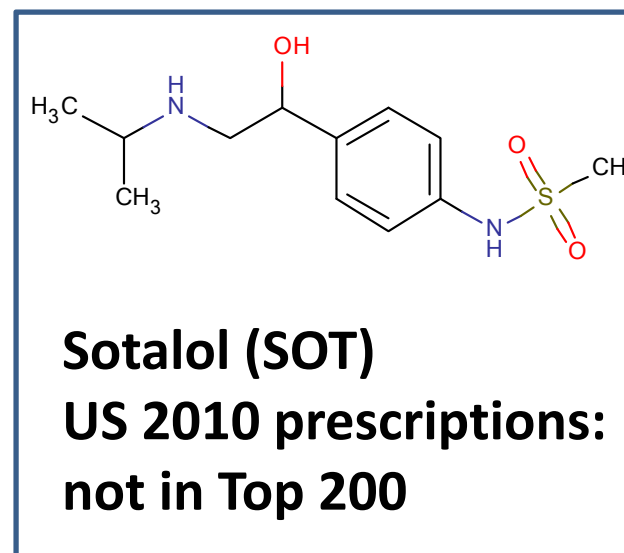
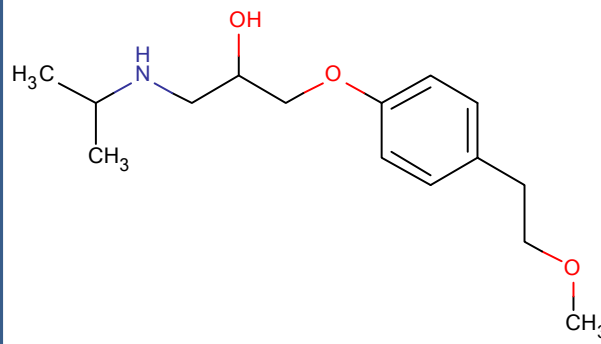
- Polyparameter models are a significant improvement; but still relatively poor as a predictive tool
- Predictive models need to capture sludge interface properties
 - question: what parameters?
 - some possible examples: CEC, EPS content

Biotransformation during nitrification

focus on three beta blockers...but also examining benzodiazepines



Metoprolol (MET)
US 2010 prescriptions:
~66.9 million (generics)



Dilute Concentrations of a Psychiatric Drug Alter Behavior of Fish from Natural Populations
T. Brodin *et al.*
Science 339, 814 (2013);
DOI: 10.1126/science.1226850

REPORTS

Dilute Concentrations of a Psychiatric Drug Alter Behavior of Fish from Natural Populations

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Environmental pollution by pharmaceuticals is increasing in aquatic environments.

OPEN ACCESS Freely available online

Diazepam and Fluoxetine Decrease the Stress Response in Zebrafish

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Abstract

The presence of pharmaceutical products in the aquatic environment has been reported in several studies. However, the impact of these drugs on living organisms is still uncharacterized. Here, we investigated the effects of acute exposure to either diazepam or fluoxetine on the stress response in *Danio rerio*. We showed that diazepam and fluoxetine inhibited the stress axis in zebrafish. Intermediate concentrations of diazepam suppressed the stress response as measured by cortisol levels, whereas fluoxetine inhibited cortisol increase at concentrations similar to those found in the environment. These data suggest that the presence of psychoactive drugs in aquatic ecosystems could cause neuroendocrine dysfunction in fish.

Abreu MSd, Koakoski G, Ferreira D, Oliveira TA, Rosa JGsd, et al. (2014) Diazepam and Fluoxetine Decrease the Stress Response in Zebrafish. *PLOS ONE* 9(7): e101322. doi:10.1371/journal.pone.0101322

BENZODIAZEPINES



at concentrations overlapping with those found in fish from River Fyris (Table 1), indicating that the treatment with low concentrations represents an environmentally relevant oxazepam contamination level. To investigate if oxazepam alters fish behavior, we quantified the behavioral traits boldness, activity, and sociality (9) of perch individuals before and after they were exposed to either of the two chosen concentrations. These referred to as perch individuals (11) for being both equally important and individuals responses (11) created at (11) sured (11) video (9), (11) : bo (11) 500 (11) f (11) ELSEVIER

PLOS ONE



β-Blockers as Endocrine Disruptors: The Potential Effects of Human β-Blockers on Aquatic Organisms

ANDREY MASSARSKY, VANCE L. TRUDEAU, AND THOMAS W. MOON*
Department of Biology and Centre for Advanced Research in Environmental Genomics, University of Ottawa, Ottawa, Ontario, Canada

β-Adrenergic blockers or β-blockers have been used therapeutically since the late 1960s. The global Aquatic Toxicology 106–107 (2012), 48–55

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Aquatic Toxicology

journal homepage: www.elsevier.com/locate/aquatox



Venlafaxine and atenolol disrupt epinephrine-stimulated glucose production in rainbow trout hepatocytes

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ABSTRACT

The beta-blocker atenolol (ATEN), and the serotonin reuptake inhibitor venlafaxine (VEN) are found in these pharmaceuticals. They affect stress response in rainbow trout hepatocytes. Venlafaxine (VEN) and atenolol (ATEN) were found to inhibit epinephrine-stimulated glucose production in rainbow trout hepatocytes. The inhibition was observed at concentrations of 10⁻⁶ M and 10⁻⁵ M, respectively. The inhibition of glucose production was not observed in the presence of the beta-blocker carvedilol. The inhibition of glucose production by venlafaxine and atenolol was not observed in the presence of the beta-blocker carvedilol. Taken together, these results suggest that venlafaxine and atenolol may disrupt glucose production in rainbow trout hepatocytes.

Keywords:
Environmental pharmaceuticals
SSNR
Beta-blocker
β-Adrenergic signaling
Cortisol
Stress response

Methods

Approved by the Ethics Committee of Universidade de Passo Fundo. Protocol #7/2013-CEUA and Conselho de Controle de Experimentação de Universidade de Passo Fundo.

1. Introduction

A diverse range of pharmaceutical compounds has been detected in municipal wastewater effluent (MWW), agricultural run-off, surface waters and even drinking water in Canada throughout the world (Metcalf et al., 2003; Servos et al., 2009; Richardson et al., 2010). Although these compounds have been detected at relatively low concentrations, their pharmacological properties and the fact that aquatic organisms, including fish, are being continually exposed to them, including the potential to cause adverse effects, have led to growing concern about the impact of pharmaceuticals on the environment.

Cardiac Pharmacology
BETA BLOCKERS



Overview of Batch Experiments

biomass from nitrification enrichment SBR

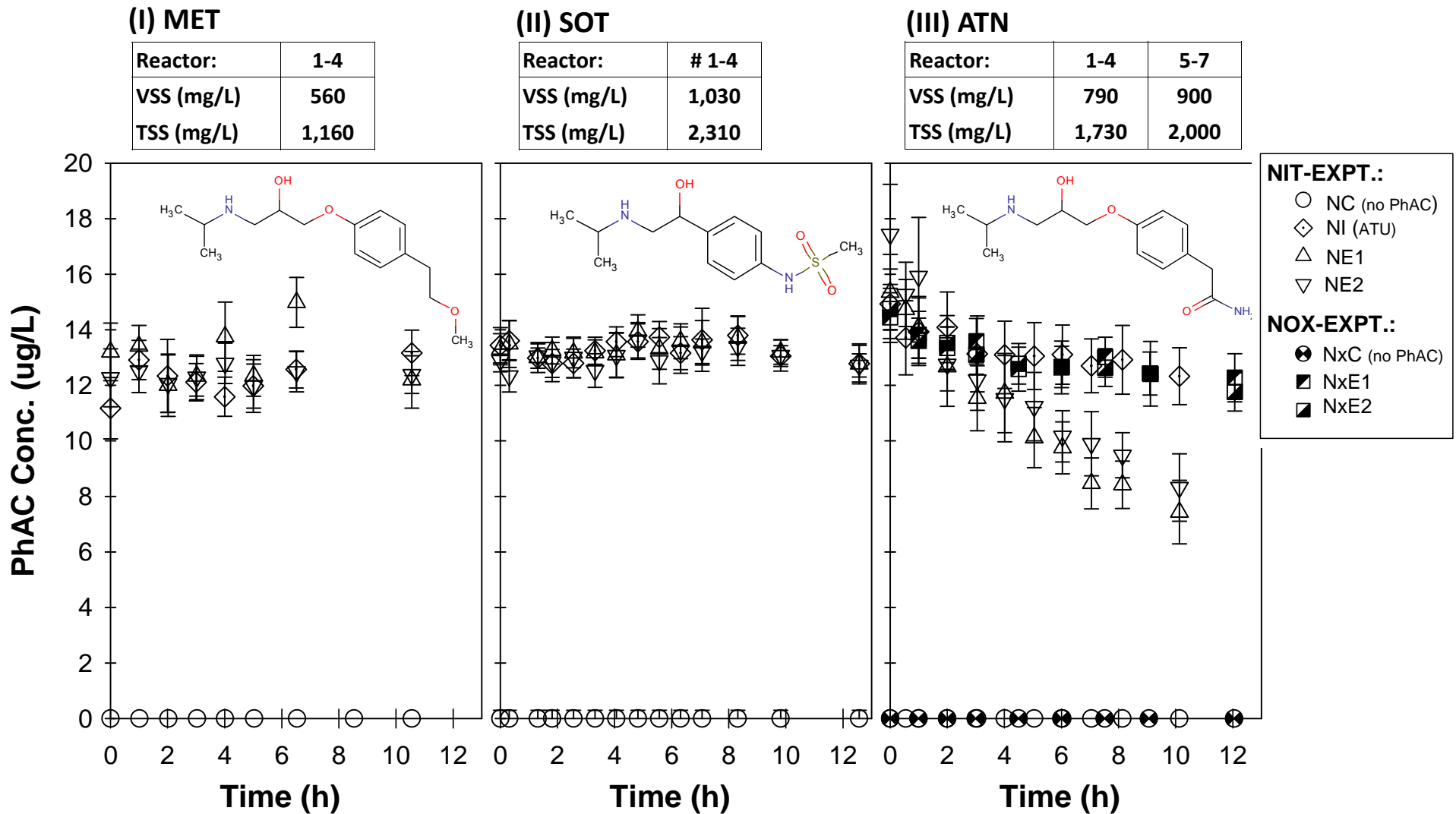
- Analyses:
 - NH_3 , NO_2^- , NO_3^- (IC-COND)
 - PhAC (LC-FLD)
 - Solids (TSS/VSS)
 - Biomass (qPCR & community profiling)



	Biodegradation during Nitrification				Biodeg during Nitrite Oxidation		
	NC (Nit. Control)	NI (Nit. Inhib. Control)	NE1 (Expt)	NE2 (Expt)	NxC (Nit. Ox. Control)	NxE1 (Nit.Ox. Expt)	NxE2 (Nit.Ox. Expt)
Biomass	✓	✓	✓	✓	✓	✓	✓
Nitrite-N	✓	✓	✓	✓	✓	✓	✓
Nutrients	✓	✓	✓	✓	✓	✓	✓
PhAC		✓	✓	✓		✓	✓
ATU		✓			✓	✓	✓

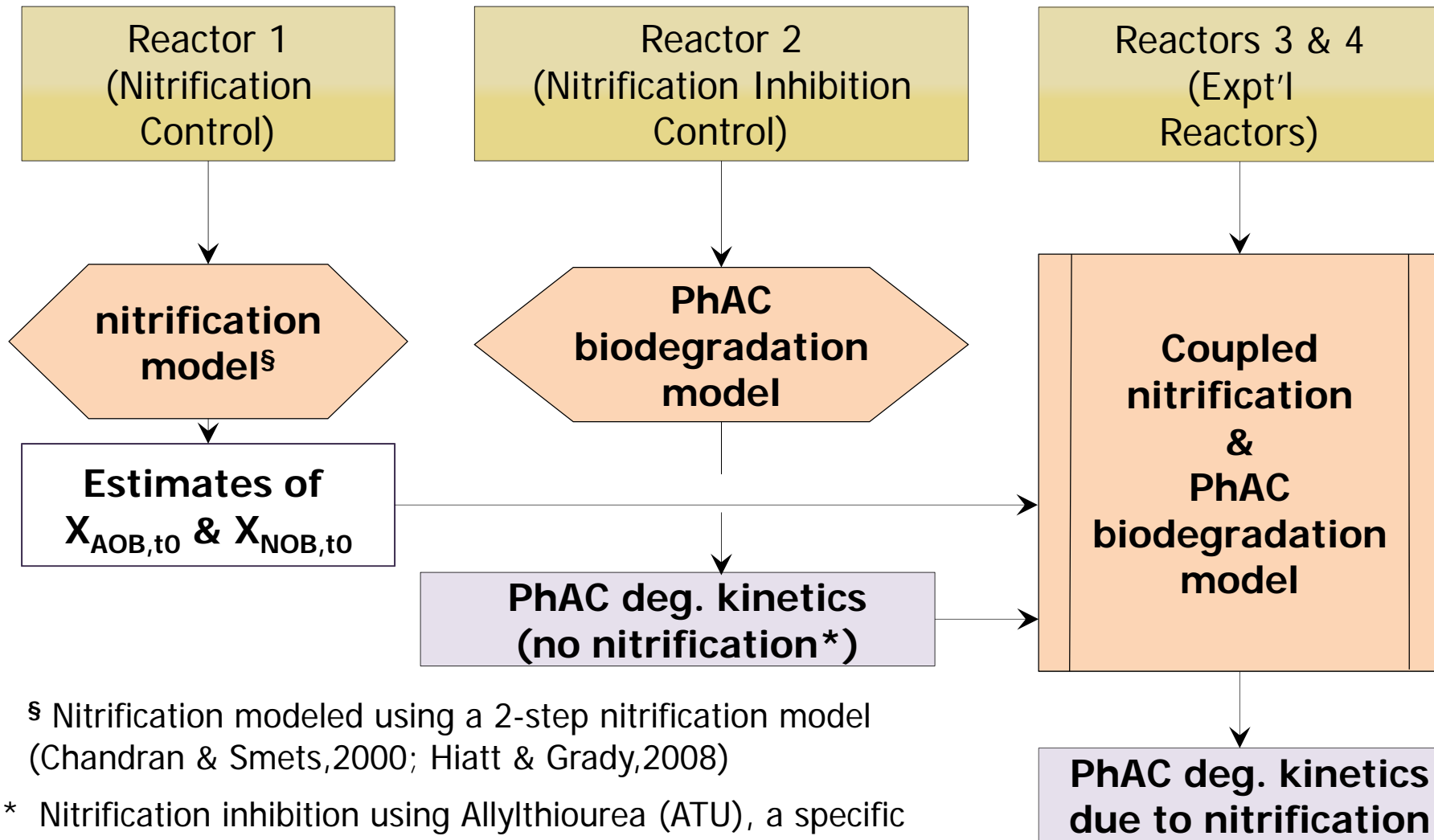
Results: ATN observed to degrade

appears linked to ammonia oxidation



Modeling Framework

Experimental data



[§] Nitrification modeled using a 2-step nitrification model (Chandran & Smets, 2000; Hiatt & Grady, 2008)

* Nitrification inhibition using Allylthiourea (ATU), a specific AOB inhibitor (Ginestet et. al., 1998)

Cometabolism Model

- Assumptions
 - due to cometabolic biodegradation by HET or AOB
 - data suggest NOB are not involved
- Functional requirements
 - PhAC biodegradation model should be adaptable for ASM framework

$$q_c = (T_c^g q_g + k_c) \frac{S_c}{(K_{SC} + S_c)} \text{ integrated model for cometabolism (Criddle, 1993)}$$



- *Modify to fit into ASM framework (assume: $S_c \ll K_c$)*

$$\frac{dS_{PhAC}}{dt} = - \left\{ \left[\frac{[T_{PhAC-AOB} \mu_{AOB}] + [k_{PhAC-AOB}]}{[T_{PhAC-HET} \mu_{HET}] + [k_{PhAC-HET}]} \right] X_{AOB} \right\} + \left\{ S_{PhAC} \right\}$$

Cometabolic Process Based (CPB) Model for PhAC biodegradation

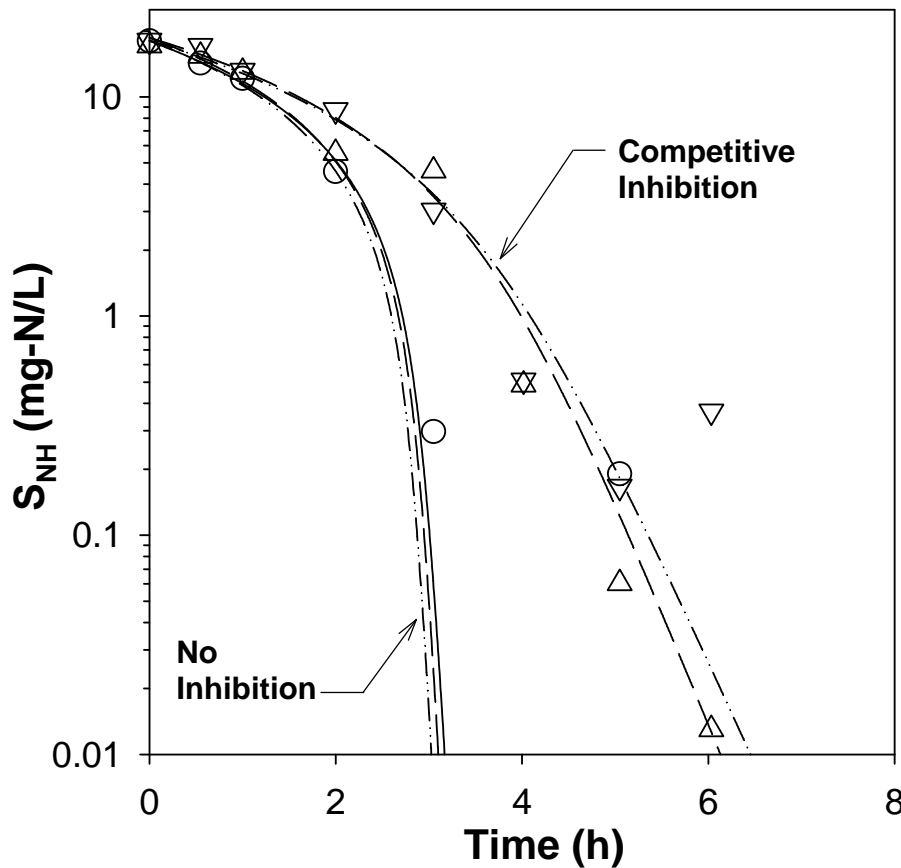


- *For batch experiments assume: $dX_{HET}/dt \sim 0 \Rightarrow X_{HET} = \text{constant}$*

$$\frac{dS_{PhAC}}{dt} = - \left\{ \left[\frac{[T_{PhAC-AOB} \mu_{AOB}] + [k_{PhAC-AOB}]}{[\alpha_{PhAC-HET} X_{HET}]} \right] X_{AOB} \right\} + \left\{ S_{PhAC} \right\}$$

Nitrification Model

PhAC inhibition of ammonia oxidation



Experimental data

- Reactor 1
- △ Reactor 2
- ▽ Reactor 3

CPB model

- Reactor 1
- - - Reactor 3
- · - · - Reactor 4

~~$$\frac{dX_{AOB}}{dt} = \left[\mu_{MAX, AOB} \left(\frac{S_{NH}}{K_{NH} + S_{NH}} \right) - b_{AOB} \right] X_{AOB}$$~~

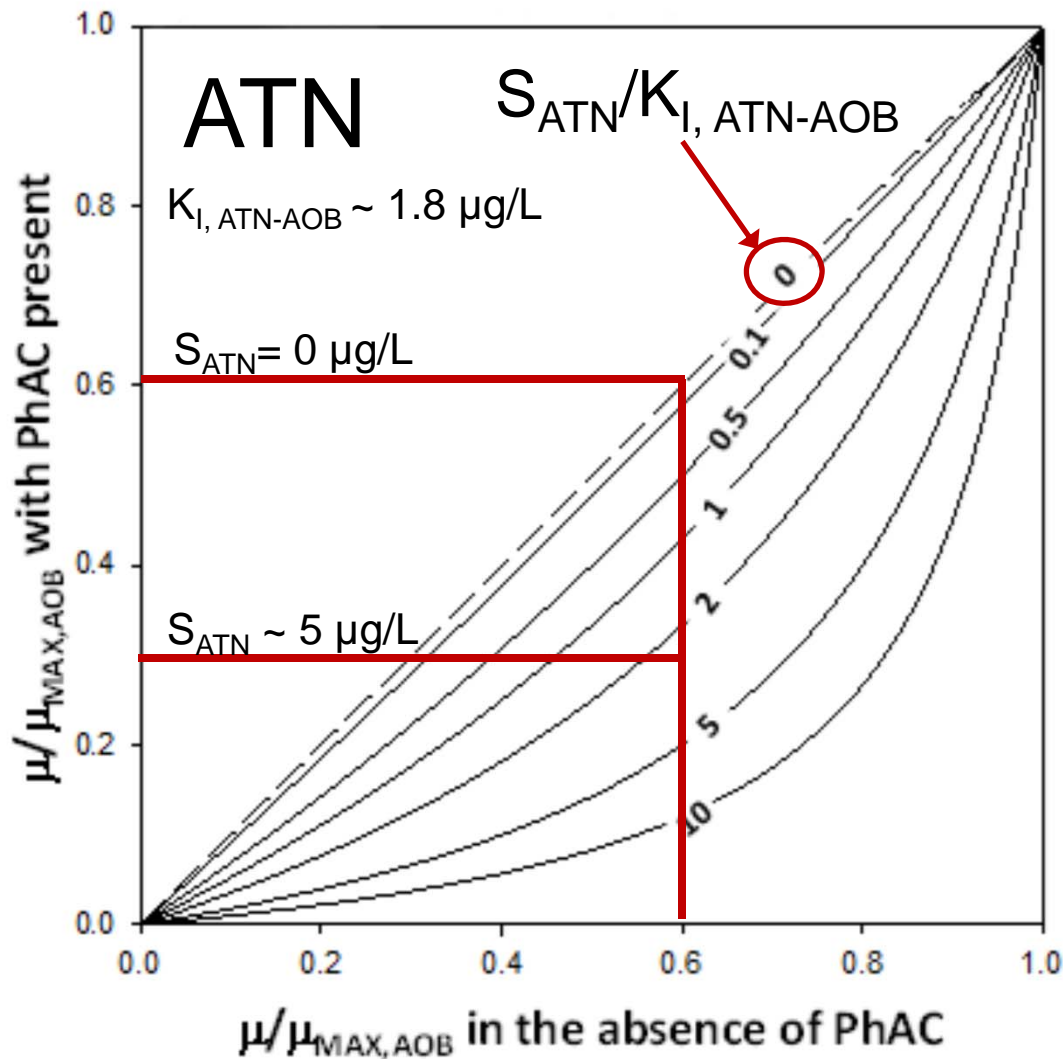
$$\frac{dX_{AOB}}{dt} = \left[\mu_{MAX, AOB} \left(\frac{S_{NH}}{K_{NH} \left(1 + \frac{S_{ATN}}{K_{I, ATN-AOB}} \right) + S_{NH}} \right) - b_{AOB} \right] X_{AOB}$$

$K_{I, ATN-AOB}$ 1.84 ± 0.39 $\mu\text{g/L}$

Goodness of fit comparison		
Model	SSE	AIC _c
No Inhibition	63.6	30.6
Competitive Inhibition	15.5	-0.4

Inhibition Implications

for atenolol



- Competitive inhibition may influence nitrification processes – needs more research
- Implication(s) for plants likely more muted
- But, competitive inhibition effects can be additive – how many PhACs exert this effect?

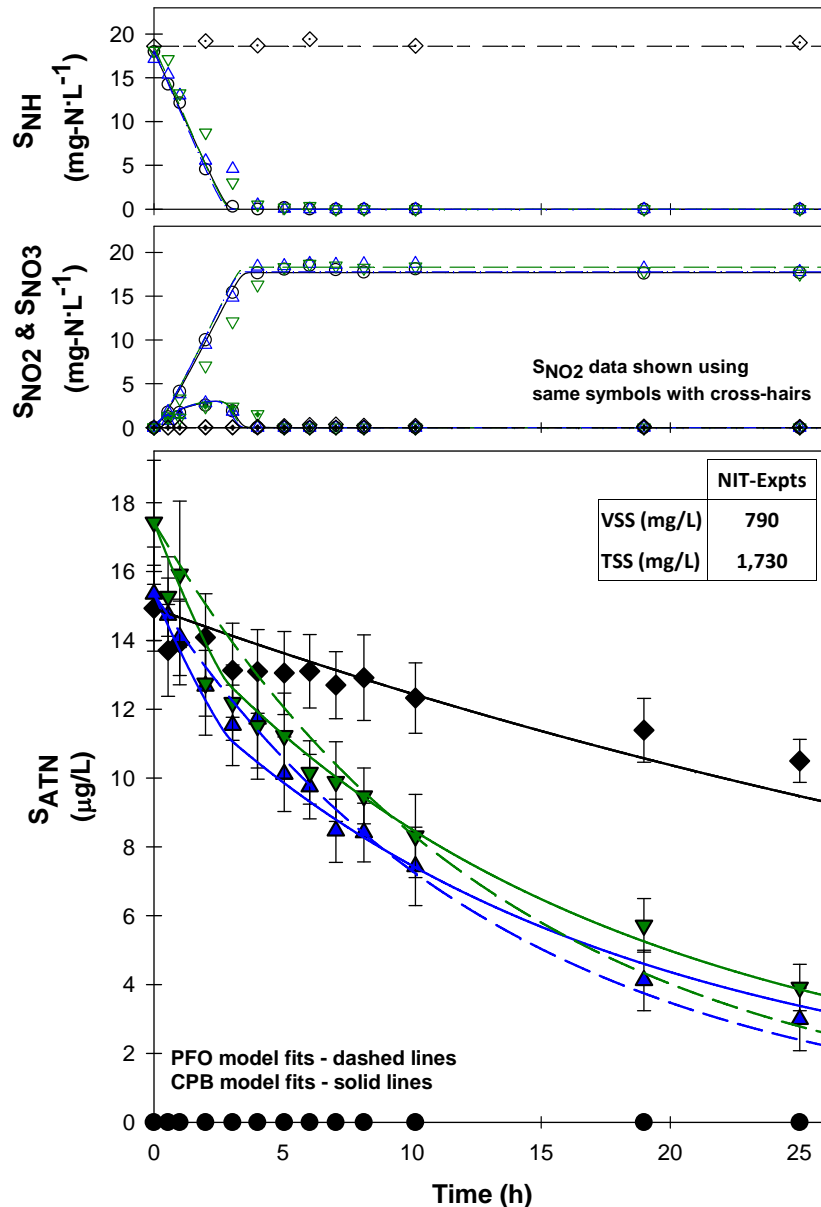
Plant influent	2.3 $\mu\text{g/L}$
Primary effluent	1.2-2.2 $\mu\text{g/L}$
Plant effluent	0.6-1.7 $\mu\text{g/L}$

Ternes et al., 2007, Lee et al., 2007

Cometabolic Process Based Model



and comparison to pseudo-first order (PFO) model



$$\frac{dS_{PhAC}}{dt} = - \left\{ \left[\left[T_{PhAC-AOB} \mu_{AOB} \right] + \left[k_{PhAC-AOB} \right] X_{AOB} \right] + \left[\alpha_{PhAC-HET} \right] X_{HET} \right\} S_{PhAC}$$

CPB model coefficients

$T_{ATN-AOB}$	71.5 ± 22.7	L.g-COD ⁻¹
$k_{ATN-AOB}$	16.1 ± 5.6	L.g-COD ⁻¹ .d ⁻¹
$\alpha_{ATN-HET}$	22.3 ± 4.4	L.g-COD ⁻¹ .d ⁻¹

Compared to 14.07
using PFO model

Sum of Square Errors (SSE)

Reactor 2	4.68
Reactors 3 & 4	4.95

Small Sample AIC for expt'l data

Best PFO model	-8.24
CPB model	-27.76

Parameter sensitivity

- Use elasticities to evaluate sensitivity of CPB model coefficients to AOB & NOB biokinetics

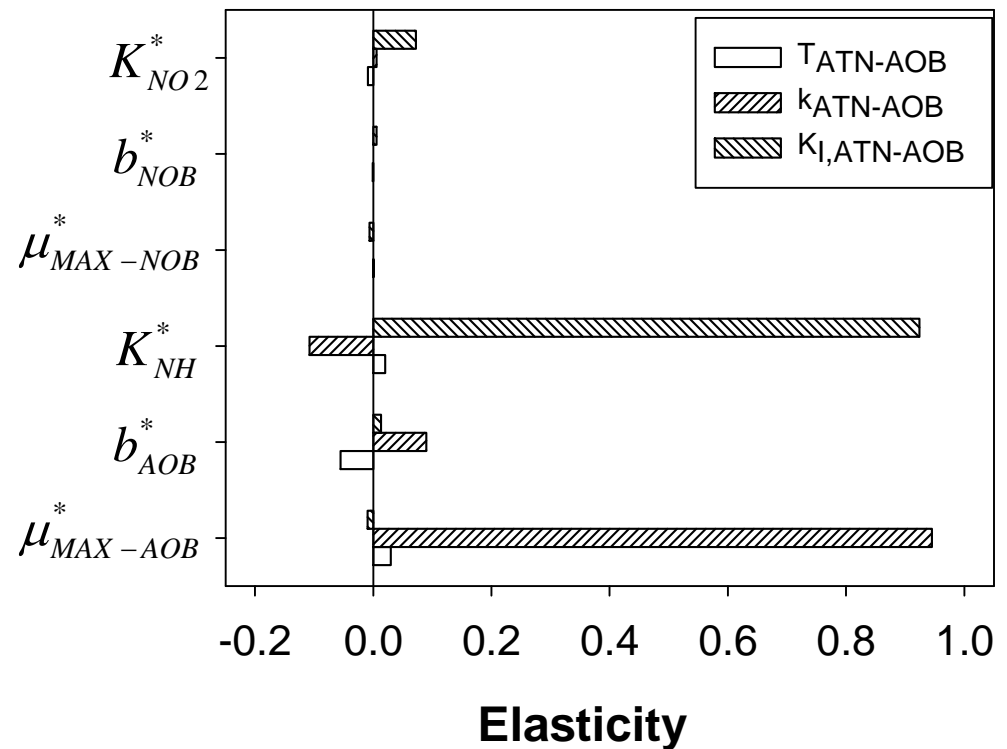
Elasticity ($\epsilon_{X/Y}$): fractional change in output (y) given a 1% change in input (x)

$$\epsilon_{X/Y} = \frac{\partial Y / Y}{\partial X / X}$$

Apply **dimensionless elasticity** (i.e., elasticity about the mean values from Monte Carlo simulations) (Sankarsubramaniam et al., 2001)

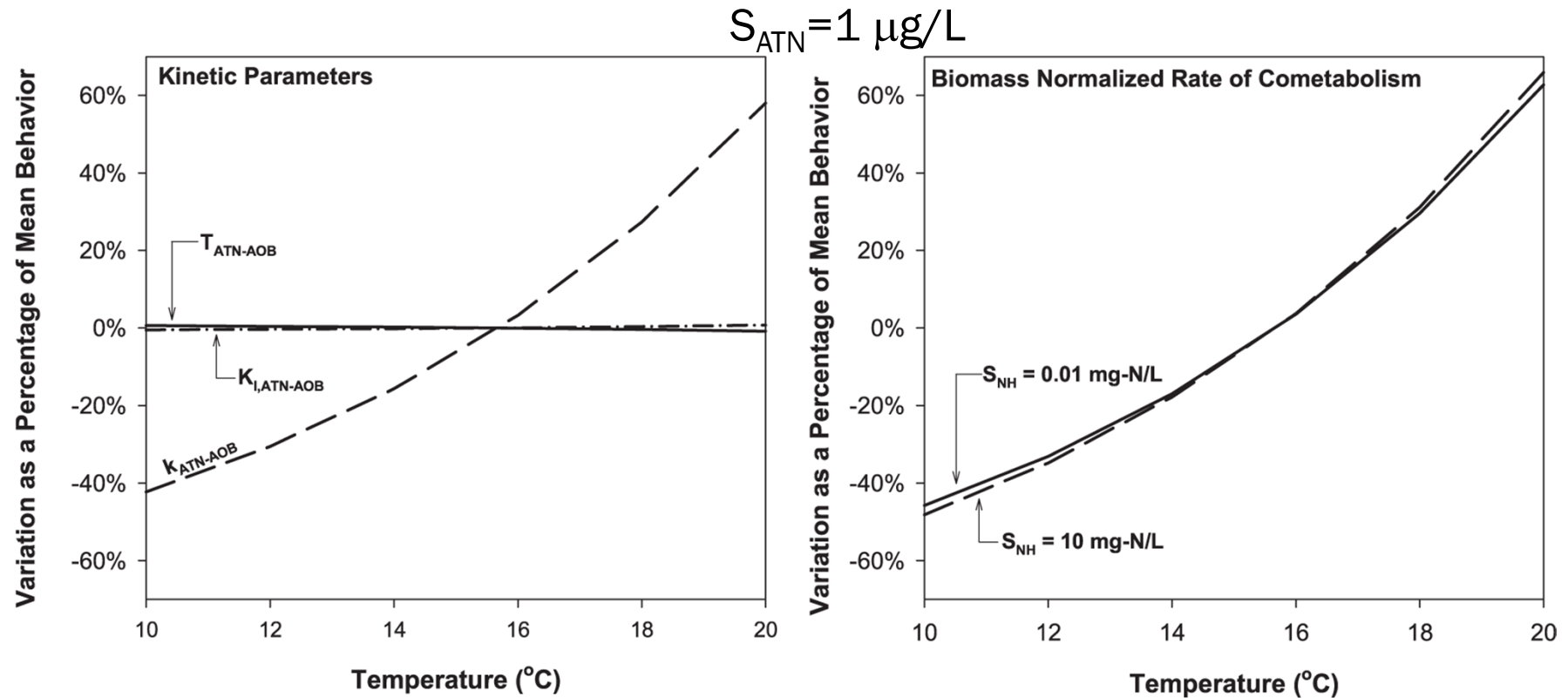
- No assumptions related to residuals behavior
- Derived using chain-rule

Use ordinary least squares regression



Implications of sensitivity

an example with temperature



- Limited influence of ammonia concentration on variation of biodegradation rate
- Temperature sensitivity of $k_{ATN-AOB}$ is due to sensitivity to $\mu_{max-AOB}$

Summary

- Atenolol and Naproxen are cometabolized by ammonia oxidizing bacteria
- Atenolol (and other PhACs) observed to competitively inhibit ammonia oxidation
- Degradation very likely results from fortuitous interactions with available/expressed ammonia monooxygenase
- Cometabolism was modeled using a new approach – cometabolic process based model (CPB)
- Model parameters are relatively insensitive to nitrification biokinetic parameters – T is insensitive, k sensitive to maximum specific growth rate
- CPB was developed to be integrated within the ASM framework, and can be readily expanded to include mechanistic description of degradation by heterotrophs.